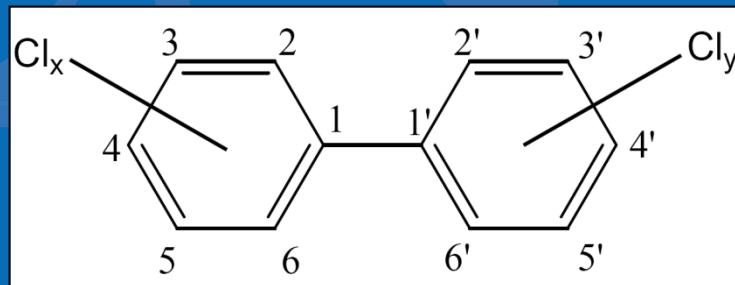


# Introduction to EPA's Human Health Risk Assessment Practices for Chemical Mixtures

*Glenn E. Rice, ScD*

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*Assessment Managers:  
Geniece M. Lehmann, PhD  
Krista Christensen, PhD*



# Disclaimer

- The views expressed in this presentation are those of the author and do not necessarily reflect the views or policies of the U.S. EPA.
- I have no conflicts of interest to disclose.

# Public Science Meeting on PCB Mixture Assessment Methods

- **Introduction to EPA's human health risk assessment practices for chemical mixtures**
  - *Glenn Rice, U.S. EPA*
- **Mixtures modeling: methods considered for the assessment of PCBs**
  - *Laura Carlson and Jeff Gift, U.S. EPA*
- **Methods for estimating relative potency values**
  - *Grace Patlewicz, U.S. EPA*
- **Overview of the Mixture Similarity Tool (MiST)**
  - *Graham Glen and Joanne Trgovcich, ICF*

# Talk Outline

1. Introduction
2. Component Methods
3. Whole Mixture Methods
4. Sufficient Similarity Methods

# Chemical Mixtures

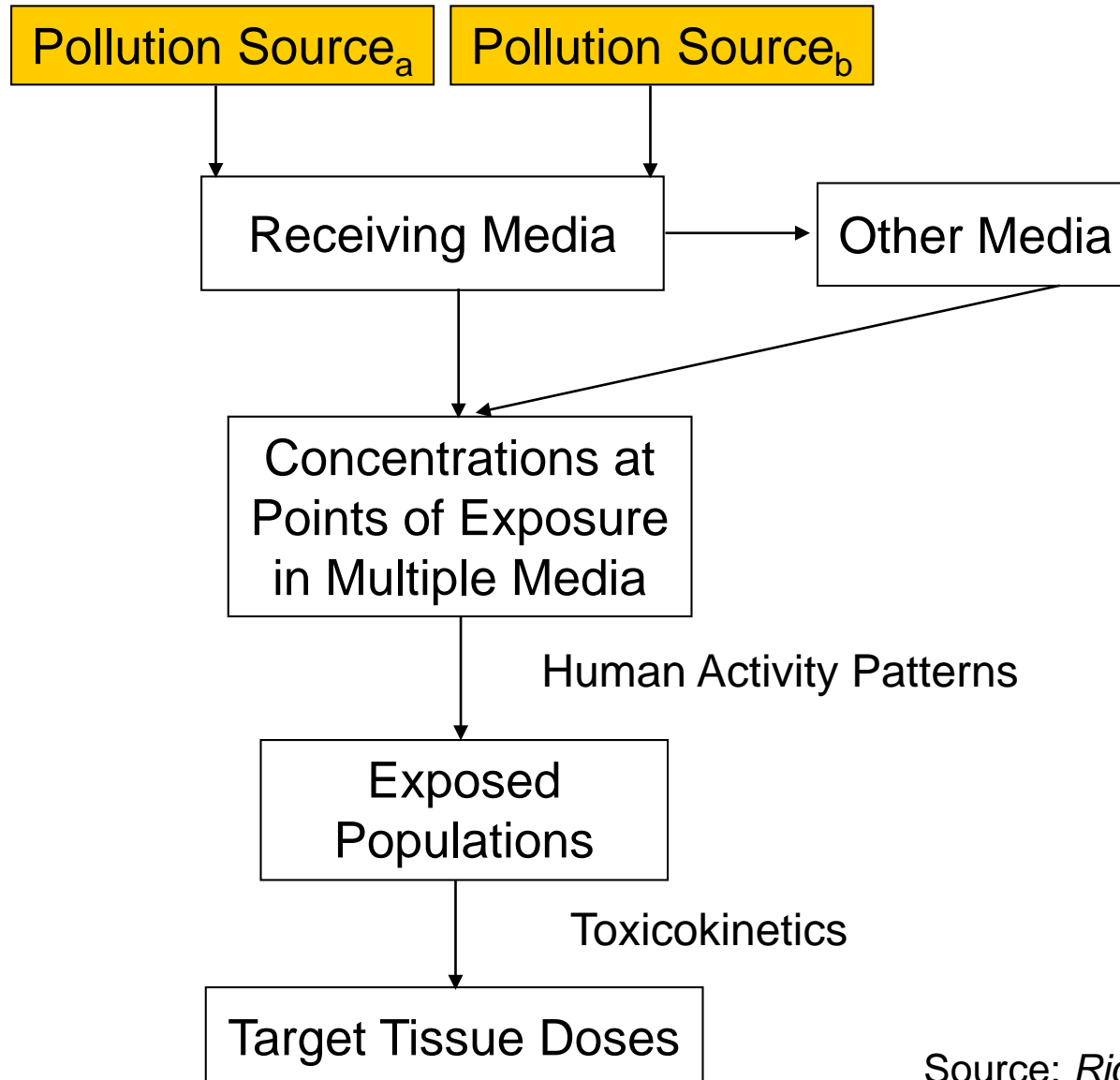
*Any combination of two or more chemical substances regardless of source or of spatial or temporal proximity*

- Combinations range from a few to 100's of chemicals
- Some components of complex mixtures may not be identified chemically

US EPA (1986) Chemical Mixtures Risk Assessment Guidelines

US EPA (2000) Supplementary Chemical Mixtures Guidance

# Mixture Exposure Assessment Conceptual Model



# Mixtures Fate and Transport

- Environmental mixtures can change over time

## ***Differential fate of mixture components***

### ***Transport***

through individual compartments

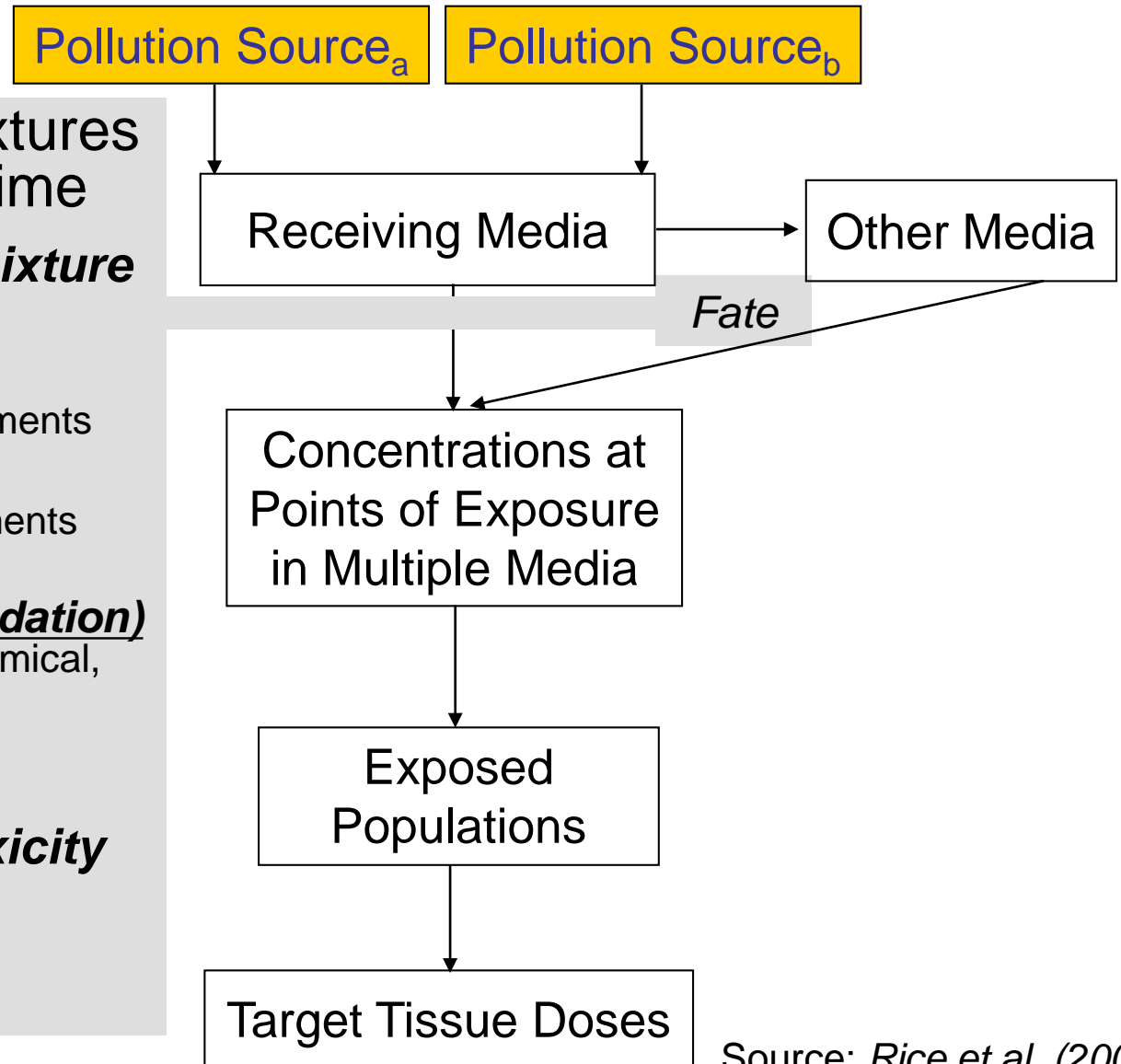
### ***Partitioning***

transfer between compartments (abiotic and biotic)

### ***Transformation (degradation)***

mediated by biological, chemical, physical agents

***Changes can affect composition and toxicity of mixture***



Source: Rice et al. (2008)

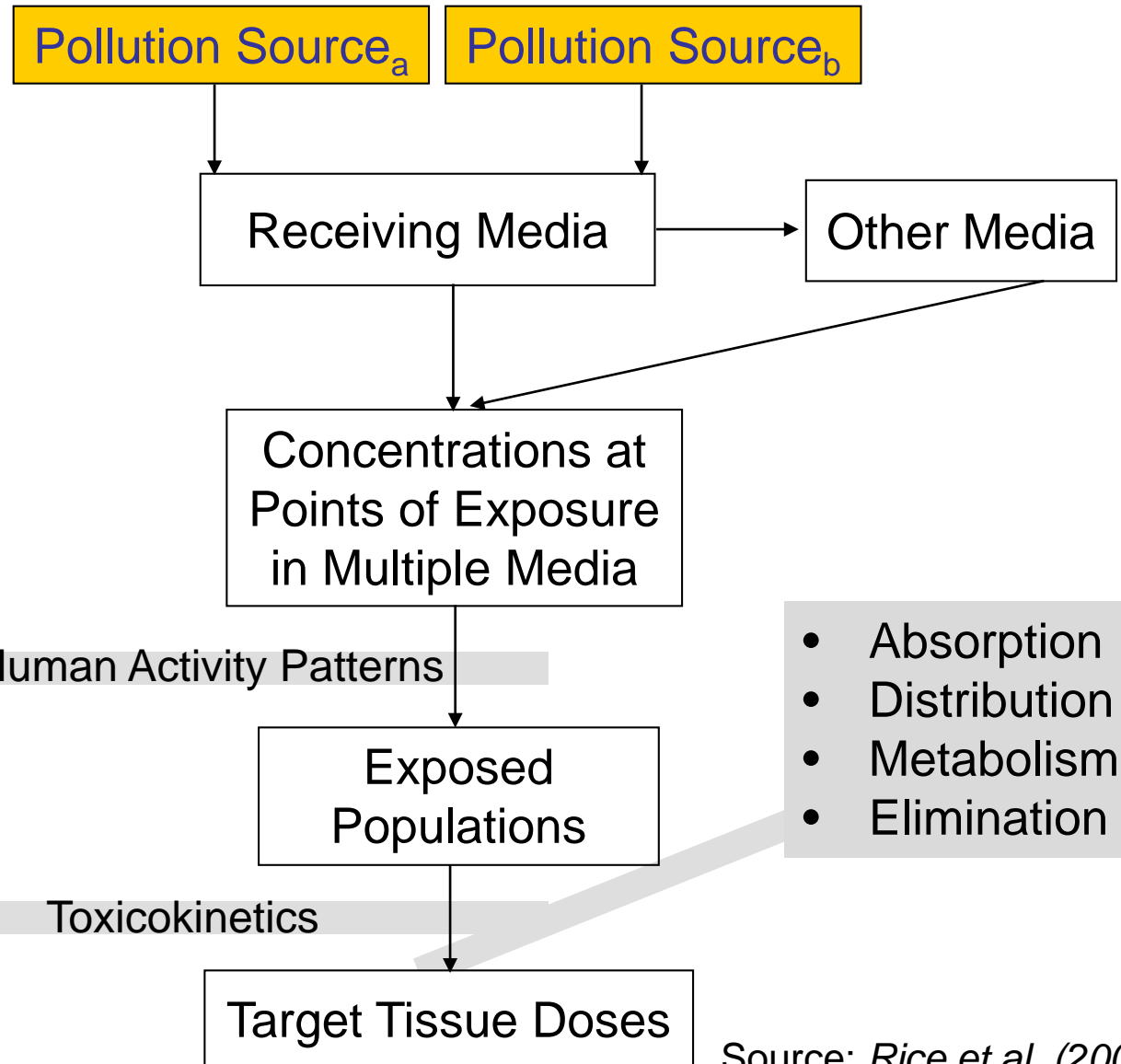
# Mixture Exposures

Humans exposed concurrently and sequentially to many chemicals

- various routes of exposure
- over varying periods of time

*Primary exposure routes:*

- ingestion
- dermal absorption
- inhalation

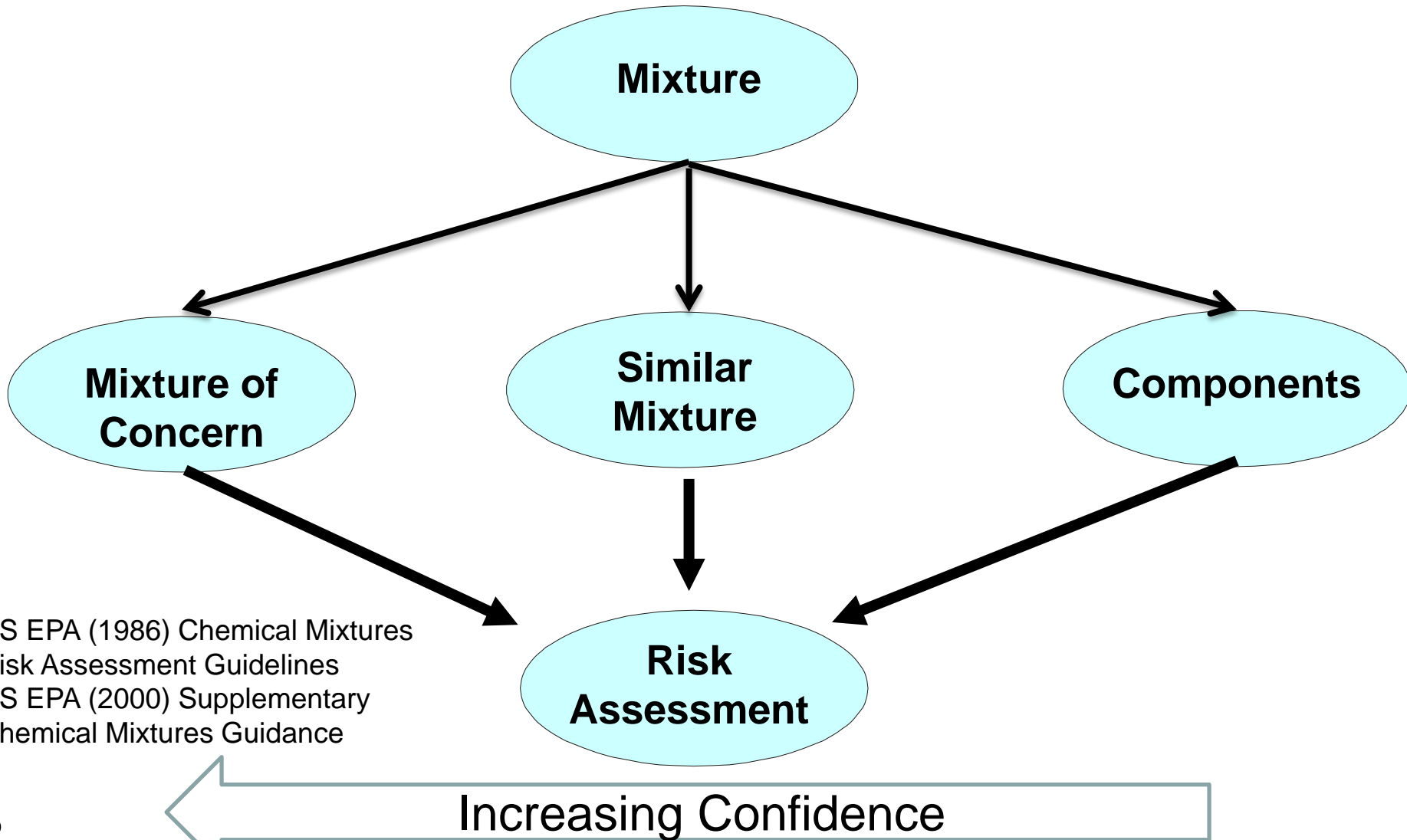


Source: Rice et al. (2008)



# Chemical Mixtures Health Risk Assessment: Approaches

Choice of approach is data-driven



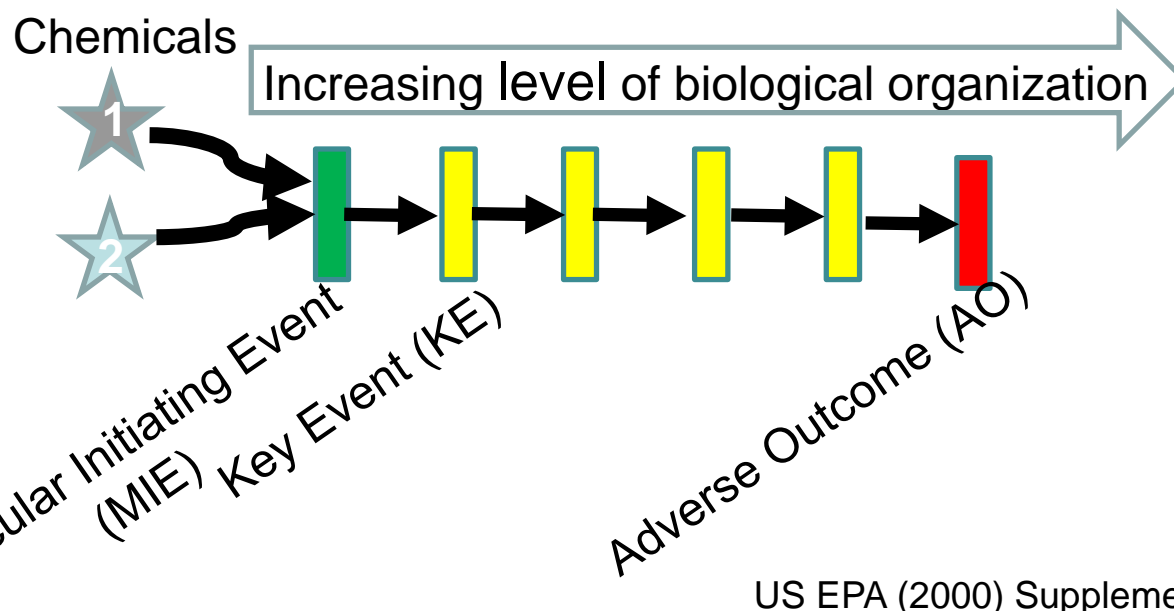
Generally, based on 1 of 3 assumptions regarding joint toxic action

1. Simple Similar Action
2. Simple Dissimilar Action
3. Toxicological Interaction

# Additive Joint Toxic Action: Simple Similar Action

Dose addition: hazard index (HI), toxicity equivalence factors (TEFs), relative potency factors (RPFs)

- Addition of component doses, scaled for relative toxicity
- Assumes components affect same pathway of toxicity (i.e., common MOA)

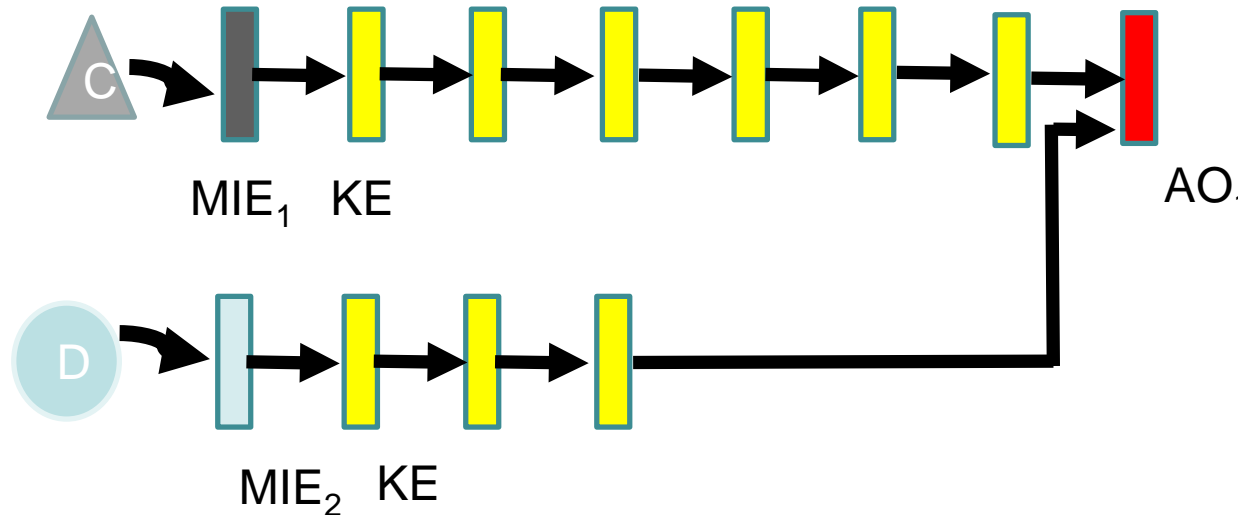


US EPA (2000) Supplementary Chemical Mixtures Guidance

Mixture of two chemicals, Chemical 1 and Chemical 2, act as toxicodynamic clones, affect same adverse outcome through same mode of action; doses add at the MIE in this hypothetical toxicity pathway

# Simple Dissimilar Action

- Response addition—cancer risk sums
  - Addition of component risks
  - Assumes toxicological and statistical independence
- Effects addition—cumulative effects
  - Addition of biological responses across components
  - Assumes toxicological similarity across components



Mixture of 2
   
 toxicologically
   
 independent
   
 chemicals affect
   
 same adverse
   
 outcome through
   
 different and
   
 independent
   
 pathways

# Toxicological Interactions

- Toxicological interactions are defined here as any toxic response that is greater than or less than those observed under the specified type of additivity, including new responses (not observed when chemicals dosed individually)
- Interaction effects
  - Types of Interactions
    - Chemical-Chemical
    - Toxicokinetic
    - Toxicodynamic
  - Many applicable terms (e.g., inhibition, masking, etc.)
  - Most common terms refer to descriptors that are:
    - greater than additive (i.e., synergism)
    - less than additive (i.e., antagonism)
- Interaction-Based Hazard Index Method

## Dose Addition: Relative Potency Factors (RPFs) Generalized Index Chemical Method

Formula for estimating the Index Chemical Equivalent Dose (ICED). The product of the RPF and the dose of the individual chemical is summed to express the mixture dose for  $n$  chemicals in terms of the index chemical:

$$ICED = \sum_{i=1}^n [RPF_i \times D_i]$$

where,

- ICED = mixture dose expressed as dose of the index chemical
- $D_i$  = dose of the  $i^{\text{th}}$  mixture component ( $i = 1, \dots, n$ ), and
- $RPF_i$  = relative potency factor is a toxicity proportionality constant relative to the index chemical for the  $i^{\text{th}}$  mixture component ( $i = 1, \dots, n$ ).

# Dose Addition: Relative Potency Factors (RPFs) Generalized Index Chemical Method

## RPF Method for 2 Chemical Mixture

$$R_m = f_1(D_1 + RPF_2 D_2) = f_1(\text{ICED})$$

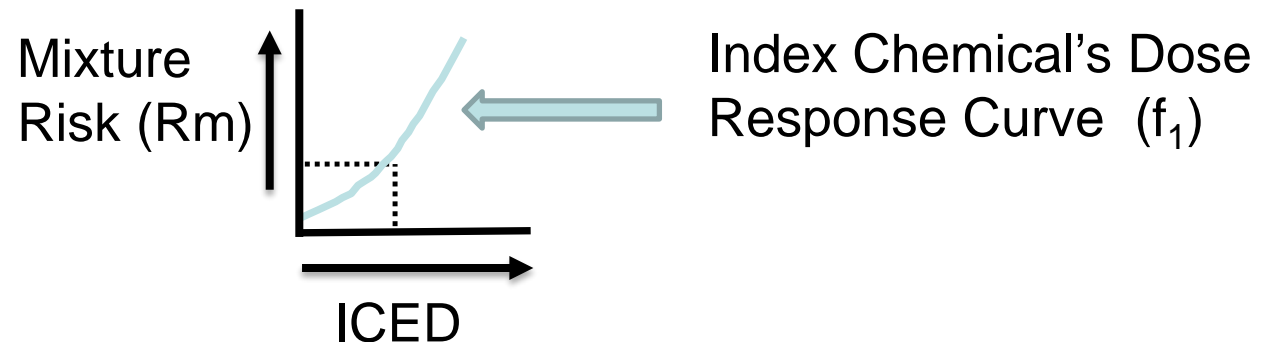
Where:

$RPF_i$  Scales the dose (D) of chemical 2 for its relative potency to index chemical (Chemical 1),

$R_m$  predicted mixture risk

ICED Index Chemical Equivalent Dose

f Index Chemical Dose-Response Function



# Methods to Calculate RPFs

For mixture components, chemical  $i$  and index chemical, the Relative Potency Factor ( $RPF_i$ ) may be estimated as:

- 1) the ratio of equally toxic doses of the 2 chemicals, e.g.,

$$RPF_i = \frac{ED_x(\textit{Index Chemical})}{ED_x(\textit{Chemical}_i)}$$

$ED_x$  = The “Effective Dose” at which an  $x\%$  response is observed.

- 2) the ratio of potency factors of the 2 chemicals, e.g.,

$$RPF_i = \frac{\textit{Dose Coefficient}(\textit{Chemical}_i)}{\textit{Dose Coefficient}(\textit{Index Chemical})}$$



# Comparison of TEFs and RPFs

## Toxicity Equivalence Factor

### Specific Type of RPF

All health endpoints

All routes

All timeframes of exposure

Encompasses all doses

Implies more abundant data

Implies greater certainty  
about mechanism of action

One TEF set for all scenarios

## Relative Potency Factor

### Generalized Case

May be limited

May be limited

May be limited

May be limited to specific dose range

May be based on lower quality/  
fewer data are available

Assumes similar mode of action  
May be more accurate because  
application can be constrained  
given available data

Can generate different RPF sets for  
various scenarios

- “Whole mixtures” typically represent the combination of chemicals in the exposure being assessed.
  - Operationally defined in some situations
- The composition of the mixture including the component chemicals and their proportions might be fully known, partially known, or unknown.

# Whole Mixture Methods

- Risk assessors generally have more confidence in assessments based on whole mixture methods than those based on component methods
- Generally, fewer data on whole mixtures
- Applicability concerns
  - composition of tested mixture may differ from environmental mixture of concern
  - complications associated with measuring mixture, and preparing mixture for toxicological testing environmental mixtures (e.g., collecting, concentrating and storing)

# Procedure to Derive Whole Mixture Health Risk Values

1. Data collection and evaluation of mixture that was tested
  - Epidemiology/human data
  - In vivo toxicology data
  - In vitro toxicology data
2. Evaluate stability of the mixture that was tested
  - Variability in components and their relative proportions
    - Across sources
    - Over time within a medium
    - Across media (e.g., uptake and retention of toxic compounds in food web)
3. Is the mixture on which health effects data are available sufficiently similar to mixture of concern (i.e., mixture encountered in the environment)?

# Procedure to Derive Whole Mixture Health Risk Values (Part 2)

4. Conduct dose-response assessment
  - Use single chemical procedures (e.g., RfD, slope factors)
  
5. Characterize uncertainties
  - Relevance of observed health effects in the study of tested mixture to those anticipated through environmental exposures

# Whole Mixture Reference Dose (RfD<sub>m</sub>)

$$RfD_m = \frac{NOAEL, LOAEL \text{ or } BMDL_X}{UF_m}$$

Where:

RfD = Reference dose

NOAEL/LOAEL = No/lowest-observed-adverse-effect level

BMDL = Lower 95% confidence limit on an X% effective dose (e.g., ED<sub>10</sub>)

UF<sub>m</sub> = Uncertainty factors for the mixture (e.g., interspecies, intraspecies, exposure duration, NOAEL to LOAEL, database deficiencies)

NOAEL, LOAEL or BMDL typically from experimental toxicity data on complex mixture dose-response

Reference Dose “an estimate of a daily exposure to the human population (including sensitive subgroups) that is likely to be without an appreciable risk of deleterious effects during a lifetime.”

# Mixture Case Study: Technical and Weathered Toxaphene

- Manufacture of insecticide, piscicide, and acaricide begins mid-1940s
- Complex mixture of hundreds of chlorinated terpenes (ATSDR, 2014)
- U.S. EPA cancelled registration for most uses in 1982 and all registered uses in 1990 (ATSDR, 2014; US EPA, 2018 )
- Most European Nations ban toxaphene during 1980s (Barbini, 2007)
- Following environmental release, toxaphene congeners undergo differential transformation and degradation via abiotic and biotic (e.g., soil microbes) processes, resulting in different mixtures of persistent toxaphene congeners, commonly termed “weathered toxaphene”

ATSDR (Agency for Toxic Substances and Disease Registry). (2014). Toxicological profile for toxaphene [ATSDR Tox Profile]. Atlanta, GA: U.S. Department of Health and Human Services. <http://www.atsdr.cdc.gov/ToxProfiles/tp94.pdf>

Barbini, (2007). Determination of toxaphene residues in fish foodstuff by GC-MS. Bull Environ Contam Toxicol 79: 226-230.

<http://dx.doi.org/10.1007/s00128-007-9179-6>

US EPA (2018). Provisional Peer-Reviewed Toxicity Values for Technical Toxaphene (CASRN 8001-35-2) Weathered Toxaphene, and Toxaphene Congeners. EPA/690/R-18/002

# Derivation of a Provisional Chronic p-RfD for Technical Toxaphene

**Effect:** thyroid cytoplasmic vacuolation, male rats, 25–29 wks dietary exposure

Point of Departure: **BMDL<sub>10</sub> (HED) = 0.0092 mg/kg-day**

- UF<sub>A</sub> = 3 (10<sup>0.5</sup>)** accounts for uncertainty in characterizing toxicokinetic or toxicodynamic differences between rats and humans
- UF<sub>D</sub> = 3 (10<sup>0.5</sup>)** accounts for uncertainty: potentially more sensitive immune effects following chronic exposure
- UF<sub>H</sub> = 10** accounts for intraspecies (human-to-human) variability in susceptibility

$$RfD = \frac{BMDL_{10}}{UF_A \times UF_D \times UF_H}$$

$$9 \times 10^{-5} \text{ mg/kg-day} = 0.0092 \text{ mg/kg-day} \div 100$$



# Derivation of Screening\* Provisional Chronic p-RfD for Weathered Toxaphene

**Effect:** thyroid cytoplasmic vacuolation, male rats, 25–29 wks dietary exposure

Point of Departure: **BMDL<sub>10</sub> (HED) = 0.0092 mg/kg-day**

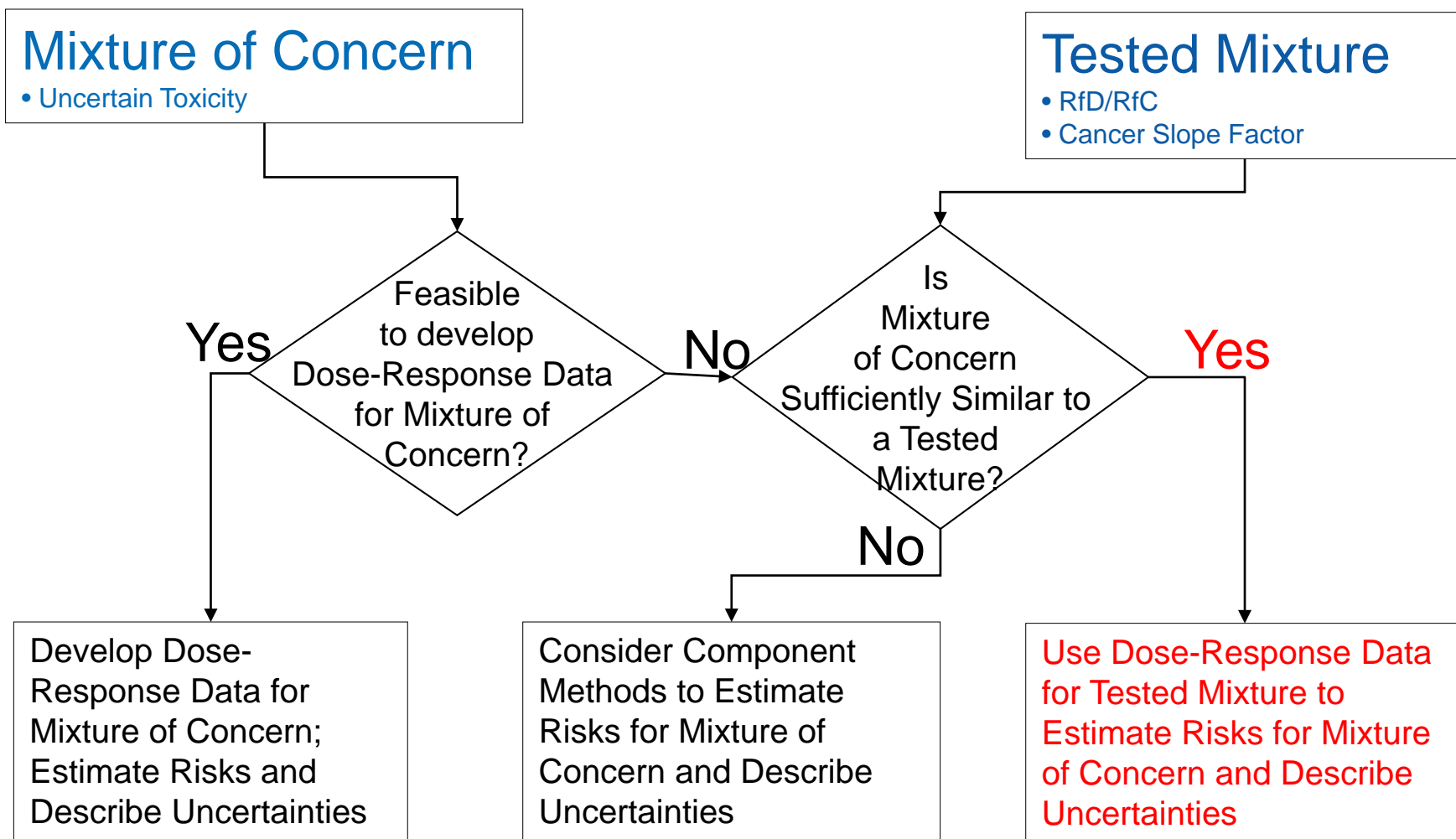
- UF<sub>A</sub> = 3 (10<sup>0.5</sup>)** accounts for uncertainty in characterizing toxicokinetic or toxicodynamic differences between rats and humans
- UF<sub>D</sub> = 10** accounts for uncertainty associated with the limited available data comparing the toxicity of technical toxaphene and weathered toxaphene
- UF<sub>H</sub> = 10** accounts for intraspecies (human-to-human) variability in susceptibility

\*EPA judged that the toxicity data for weathered toxaphene or individual toxaphene congeners are inadequate to derive noncancer provisional toxicity values. However, available information was judged adequate to be of limited use to risk assessors, and in such cases, EPA develops a “screening value.”

## EPA 2000

- If toxicity data are not available for a mixture of concern, the risk assessment can be based on surrogate toxicity information obtained from testing a *sufficiently similar* mixture
- Describes general principles, but no methods
  - A mixture is sufficiently similar to another when its components are not very different and are in about the same proportions.
  - Few differences in environmental fate, uptake, bioavailability and pharmacokinetics.
  - Expected toxicological consequences of exposure to the two mixtures are nearly identical.

# Sufficient Similarity



**Mixtures considered sufficiently similar when expected health consequence of exposure to 2 mixtures nearly identical**

**Source:** Rice, G. E., I. Eide, P. I. Feder, C. Gennings. 2018 . Assessing Human Health Risks Using Information on Whole Mixtures. Chapter 15 in Chemical Mixtures and Combined Chemical and Nonchemical Stressors: Exposure, Toxicity, Analysis, and Risk Editors C.V. Rider and J.E. Simmons. Springer International Publishing AG

# Sufficient Similarity Approaches

## Sufficient similarity methods not included in EPA (2000) Guidance

### Some Published Sufficient Similarity Approaches

- Feder et al. 2009: Uses principal component analysis
- Feder et al. 2009: Uses statistical bootstrap method
- Marshall et al. 2013: Uses equivalence testing methods
- Caitlin et al. 2018: Additional NTP tox. Testing of botanical supplement. Uses strength-of-evidence approach, empirical equivalence testing, and visual interval evaluation

Marshall et al. 2013 methods provide a reasonable and defensible approach for evaluating similarity among PCB Mixtures (**discussed further in presentation 2**)

Marshall et al. 2013 "An empirical approach to sufficient similarity: combining exposure data and mixtures toxicology data" Risk Analysis 33(9) 1582-1596.

Feder et al., 2009 "Evaluating sufficient similarity for drinking-water disinfection by-product (DBP) mixtures with bootstrap hypothesis test procedures". J Toxicol Environ Health A.;72(7):494-504.

Feder et al 2009. "Evaluating sufficient similarity for disinfection by-product (DBP) mixtures: multivariate statistical procedures." J Toxicol Environ Health A. 2009;72(7):468-81.

Caitlin 2018. How similar is similar enough? A sufficient similarity case study with Ginkgo biloba extract. Food Chem Toxicol. 118:328-339.

- Basing risk assessments on whole mixtures dose-response data considered more reliable than basing assessments on component dose-response data
- Component data approaches most often used in assessments of environmental mixtures; RPF method has been used to evaluate hazard and cancer risks for multiple mixtures
- When available (infrequent), reference doses and cancer slopes can be derived from whole mixture dose-response data and used in risk assessments
- A goal of sufficient similarity approaches is to evaluate the confidence in using dose-response data for a tested mixture in assessments of other mixtures that have either not been tested or been subjected to limited testing

- Additional approaches for evaluating similarity are needed
- Outputs from new approach methodologies (NAMs) could be used to generate hazard assessment data and dose-response data to address similarity among mixtures
  - toxicogenomics
  - proteomics
  - metabolomics
  - chemoinformatics
  - bioinformatics
  - cell-based bioactivity screening assays
- Informativeness of the NAM data for addressing similarity among mixtures will depend on relationship between the NAM endpoint/s and human disease (i.e., confidence in understanding the relationship).

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For more information, please contact an IRIS PCB assessment manager: Geniece Lehmann ([lehmann.geniece@epa.gov](mailto:lehmann.geniece@epa.gov)) or Krista Christensen ([christensen.krista@epa.gov](mailto:christensen.krista@epa.gov))